1	Herbivores, resources and risks: alternating regulation along primary
2	environmental gradients in savannas.
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22 Abstract

23 Herbivores are regulated by predation under certain environmental conditions, whereas in 24 others they are limited by forage abundance and nutritional quality. Whether top-down or 25 bottom-up regulation prevails depends both on abiotic constraints on forage availability 26 and body size, because size simultaneously affects herbivores' risk of predation and their 27 nutritional demands. Consequently, ecosystems composed of similar species can have 28 different dynamics if they differ in resource supply. We use large herbivore assemblages 29 in African savanna ecosystems to develop a framework that connects environmental 30 gradients and disturbance patterns with body size and trophic structure. This framework 31 provides a model for understanding the functioning and diversity of ecosystems in 32 general, and unifies how top-down and bottom-up mechanisms are dependent on 33 common underlying environmental gradients.

34

35 Herbivore regulation and the implications of body size

36 The global decline of large herbivores, due to human-induced landland use changes, 37 raises concerns for the long-term conservation of species whose ranges are being reduced 38 to a handful of protected areas [1]. The local extirpation of large herbivores has 39 consequences for entire ecosystems, because of their role in maintaining the diversity of 40 predators and primary producers [2]. Understanding herbivore regulation across resource 41 gradients, such as rainfall, is important for the long-term management and conservation 42 of ecosystems, especially if shifts in global climate result in a mismatch between the 43 location of protected areas and a species' preferred niche. Here, we investigate how

resource gradients simultaneously influence top-down and bottom-up processes in
ecosystems, using the large herbivore community of African savannas as a generalized
example. The model could also prove useful in understanding the relation between
disturbance, resource gradients and trophic structure in other ecosystems.

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Classic food chains represent relationships between trophic levels as linear bottom-up or top-down processes: abiotic factors such as rain determine primary production, which is consumed by herbivores, which are, in turn, consumed by carnivores. The abundance of herbivores can therefore be controlled through top-down mechanisms, such as predation [3-5], or through bottom-up constraints on primary production, such as soil fertility (Figure 1) [6-9].

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Trophic cascades in linear models of herbivore regulation (Figure 1,) involve the knockon effects of predation expressed at alternate trophic levels [10]. In the classic example, predators limit the abundance of herbivores, which releases grazing pressure on plants (the 'green world' hypothesis) [11]. In this hypothesis, the abundance of vegetation is determined largely by the availability of abiotic resources as herbivores are regulated by predators [5, 12].

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Previous research has elucidated the complexity of trophic interactions by breaking each
trophic level into more fundamental components (Figure 1). Specifically, the role of
abiotic factors, disturbances, quality and quantity of primary production, and the effect of
body size have each been shown to influence independently the distribution and

67 abundance of herbivores. Here, we show how common underlying environmental 68 gradients influence both top-down and bottom-up regulation simultaneously. Differences 69 in the relative accessibility of limiting resources can cause ecosystems with similar 70 species to have different regulatory mechanisms. In addition, the body sizes of herbivores 71 determine both their susceptibility to predators and their resource requirements. We 72 show how this generalized model accounts for observed differences in the trophic 73 functioning of the large herbivore community in savanna sites across Africa. Although 74 humans evolved in African savannas and historically affected herbivores through hunting 75 and fire, substantial landuse changes and increasing human populations have put 76 unnatural demands on these systems, which calls for a better understanding of ecosystem 77 dynamics.

78

79 **Predation: not all herbivores are affected equally**

80 The simple food-chain view of predator-prey interactions (Figure 1) ignores the fact that 81 not all carnivores can consume all herbivores, and not all herbivores are equally 82 susceptible to all carnivores. Large prey, such as buffalo (Syncerus caffer), are difficult to 83 capture and are only consumed by the largest predators, such as lion. Whereas small 84 predators can only consume small prey, large predators might consume both large and 85 small prey (Figure 2a) [13, 14]. Recent work proposes that predation has a greater impact 86 on regulating a population of small body-sized herbivores (e.g. oribi, Ourebia ourebi) 87 when the prey base of small predators is nested within that of large predators as this 88 exposes smaller herbivores to more enemies (Figure 2a) [15]. Conversely, if predators 89 specialize on particular size classes of prey (i.e. they partition the prey base), predation

90 pressure is expected to be relatively even across all body sizes, until the prey become too91 large (Figure 2b).

93	Therefore, the degree of regulation of herbivore populations by predation depends on
94	both the relative body size of predators and prey, and the way in which the available prey
95	base is partitioned by carnivores of different sizes (Figure 3a). Furthermore, the largest
96	predators (lion) are also the most dominant and actively kill smaller predators (cheetah,
97	hyena), or chase them away from their prey (Figure 3a) [16, 17].
98	
99	Recent studies provide evidence that both size-nested and size-partitioned predation
100	occurs (Figure 3b). In the Serengeti system (Tanzania) large prey, such as buffalo
101	(450kg), generally escape predation except from cooperatively hunting lions, while oribi
102	(18 kg) are eaten by many species [3, 18] illustrating size-nested predation (Figure 3b).
103	Studies in Kruger Park, Hluhluwe-iMfolozi, and Phinda (South Africa) provide evidence
104	for size-partitioned predation where large predators, such as lion, concentrate on prey
105	centered around the predators' body mass [4, 19] (but see Ref. [15]). Only prey species
106	with a body mass >1000kg, such as hippo (Hippopotamus amphibius) and white rhino
107	(Ceratotherium simum), tend to escape predator regulation in Kruger, whereas prey
108	species heavier than 150kg, such as buffalo, escape predator regulation in Serengeti. The
109	largest herbivores, such as elephant (Loxodonta africana), are too big to be captured
110	irrespective of whether predation is size-nested or size-partitioned [20, 21] (with the
111	exception of juveniles).

In summary, top-down processes are modified by the way in which carnivores partition their food niches and the degree to which larger carnivores dominate smaller carnivores. Explanations as to why there are differences in the mechanism of niche partitioning of predators in otherwise similar food chains can be found in the type of vegetation that supports herbivores and the disturbance regime, which we explore below.

118

119 **Forage quality and abundance: not all that is green is edible**

120 Geographic processes involving erosion of parent material, and rainfall, determine key 121 environmental gradients, such as soil fertility and water availability [22, 23], which 122 influence vegetation structure [7, 24, 25]. Plant structure, in turn, determines the quality 123 and quantity of digestible material available to herbivores (Figure 4a) [9, 26-30]. Primary 124 production varies along environmental gradients [31] and regulates herbivore populations 125 through classic bottom-up processes of resource limitation (Figure 4a) [8, 32-34]. The 126 quantity of primary production increases with rainfall and soil fertility, such that in the 127 absence of herbivory or fire, the largest standing biomass is found in fertile areas with 128 unlimited moisture while declining when either water or soil nutrients become limiting 129 (Figure 4b) [7, 29, 35]. Under high rainfall conditions, plants invest more resources in 130 structural support and protection against herbivory (e.g. stems, lignified tissues, 131 secondary compounds and mechanical defenses [35, 36]). As a result, the digestible 132 quality of primary production is inversely related to rainfall (Figure 4b), such that the 133 amount of energy and nutrients per unit biomass that is extractable by herbivores declines 134 as conditions become wetter [37].

135

136 Herbivory provides the most direct access to consumable energy and has evolved 137 multiple times in many unrelated taxa (e.g. molluscs, birds, mammals, insects, reptiles, 138 fish, and marsupials) from both carnivorous and detrivorous ancestors [38]. To digest 139 cellulose, ungulates use a symbiotic fermentation process in the rumen or caecum that is 140 relatively time-consuming and requires a specialized gastrointestinal tract. Small 141 ungulates like oribi and Thomson's gazelle (Gazella thomsoni) have smaller 142 gastrointestinal systems, and therefore shorter ingesta retention times [20, 21, 39], which 143 means they cannot process coarse vegetation. Furthermore, small endotherms have a 144 higher energy expenditure per unit mass. These two factors mean that small herbivores 145 have to select the most nutritious, highest energy forage (Figure 4c) [26, 28, 40, 41]. 146 Larger herbivores are relatively unconstrained by the size of their gastrointestinal tract, 147 they have longer retention times, and so can extract sufficient energy from poorer quality 148 food, providing there is sufficient quantity [40, 42] (Figure 4c).

149

In summary, plant quality and biomass are determined by both environmental gradients and plant growth form (Figure 4a). These affect small and large herbivores differently due to differing metabolic constraints. The result is that smaller herbivore populations are nutritionally limited by the quality of forage, whereas populations of larger grazers are limited by the quantity of food (Figure 4d).

155

156 The role of disturbances, facilitation and ecosystem engineers

157 Sudden shifts in primary production caused by abiotic disturbances such as fire, or

158 marked changes in consumption rates (either herbivory, predation or infection)

potentially rearrange the dynamics of an ecosystem either temporarily or semipermanently into a new state [43, 44]. Given that local densities of herbivores can be
regulated by both predation and attributes of the plant community (e.g. structure, quality
and quantity), disturbances are factors that can change the primary mechanism of
herbivore regulation and lead to nonlinear responses in abundance (Figure 5a) [22, 4547].

165

166 Reciprocal effects occur between large herbivores and primary producers (represented 167 with double arrows in Figure 5a) that can lead to grazing facilitation between species [25, 168 48, 49]. Mega-herbivores, such as white rhino or hippo, create and maintain low-biomass 169 grass swards composed of nutritious grazing-tolerant grasses, which subsequently support 170 other smaller grazers [50, 51]. The repetitive grazing of specific patches by multiple 171 species where more dung and urine are deposited might have similar consequences, 172 resulting in fertile hotspot locations where grazers consistently occur over time [52, 53]. 173 174 Reciprocal interactions alter the probability of a disturbance occurring in an ecosystem, 175 whereas interaction modifiers (as per Ref. [14]) alter the severity of a disturbance. For 176 example, the relative proportion of trees and grasses in a savanna influences its potential 177 flammability [54] because grasses (which senescence seasonally) contribute more to the 178 fuel load than do trees. Once grasses dominate the plant community, owing to 179 disturbances such as herbivory (e.g. Ref. [55]), a positive feedback between grass 180 abundance and fire frequency can arise (double arrow in Figure 5a). In addition, the 181 accumulation of grass biomass also alters the intensity of a fire (dotted line in Figure 5a),

182 which when combined with the positive feedbacks between fire frequency and grass 183 abundance, maintains an open grassland landscape, preventing tree invasion [54, 56-58]. As a result, fires can prevent invasion of grasslands by trees, which is engineered, in part, 184 185 by the grazing intensity of herbivores [59, 60]. 186 187 Disturbances such as grazing and fire can act additively in savanna systems by changing 188 the competitive balance between grazing tolerant and intolerant grasses (Figure 5b) [46] 189 and influencing the nutritional quality of the forage supporting herbivores (Figure 5c) 190 [61, 62]. So where some savannas have sufficient rainfall to support closed forests, they 191 persist as mixed grasslands owing primarily to disturbances [46, 63]. Thus, two systems 192 with similar rainfall and nutrient regimes could have different woodland-grassland 193 structures because disturbances push systems between multiple states [55, 58] and this, in 194 turn, affects the abundance of herbivores.

195

Emerging properties: top-down and bottom-up processes are not mutually exclusive

198 The separate roles of predation (Figure 3a), primary production (Figure 4a), or

199 disturbance (Figure 5a) in the regulation of herbivore populations have different

200 consequences when they are combined as opposed to when considered separately.

201 Underlying environmental and landscape gradients affect top-down and bottom-up

- 202 processes simultaneously [2] by influencing the forage quality and quantity available to
- 203 herbivores while changing their exposure to predation (Figure 6a). Thus, in Kruger
- 204 rainfall simultaneously affects predation and primary production [64]. Moreover, the

topography of the landscape (e.g. hills, catenas and rivers) affects the quality and
abundance of primary production available to herbivores [7, 22, 45], while at the same
time determines the amount of cover for a predator, and increases their hunting success
by concealing them from their prey (Figure 6b, c) [65-69]. These indirect (Figure 4a),
additive (Figure 5a), reciprocal (Figure 5a), and interaction-modifying (Figure 5a)
relationships strengthen the interdependencies between primary production, disturbances
and predation in regulating herbivores (Figure 6a).

212

213 The scale of heterogeneity determines the mode of regulation

214 Top-down and bottom-up processes are modified by the scale of both spatial and 215 temporal heterogeneity [70]. Suitable habitats can occur heterogeneously at a coarse scale 216 where large patches are separated by long distances, or as a fine-scale mosaic [71, 72]. 217 When heterogeneity is coarse, herbivores migrate long distances between suitable 218 patches. When habitats are locally heterogeneous animals move frequently between small 219 patches but do not move far. On a temporal scale, strong seasonality causes animals to 220 move between patches as phenological conditions change (such as the seasonal drying of 221 grass), whereas weak seasonality enables animals to remain in local areas. The 222 consequences of these different scales are seen in the long-distance seasonal migrations 223 of wildebeest (Connochaetes taurinus) in Serengeti, Coke's hartebeest (Alcelaphus 224 buselaphus) on the Athi plains of Kenya, or white-eared kob (Kobus kob leucotis), topi 225 (Damaliscus lunatus) and Mongalla gazelle (Gazella thomsoni albonotata) in Sudan [1]. 226 Such movements reduce the impact of predation as predators cannot follow the 227 herbivores over these long distances [73, 74] and result in bottom-up regulation [34] (Box

228 1). By contrast, fine-scale spatial and temporal heterogeneity keeps herbivores within the 229 territories of predators and results in top-down regulation. This distinction is seen by 230 comparing resident wildebeest in Pilanesberg, Hluhluwe-iMfolozi, Kruger (South 231 Africa), Ngorongoro Crater (Tanzania), and Etosha (Namibia) [4, 75-78] which are 232 predator-regulated, to migrant wildebeest in Serengeti [34] which are food-regulated 233 (Box 1). Therefore, abiotic gradients influencing the regulation of herbivores at large 234 scales through predation and nutrition are replicated at finer scales and determine how 235 animals are distributed over a landscape [79].

236

237 Abiotic gradients determine the direction of regulation

238 The availability of abiotic factors, such as soil nutrients and rainfall, determines the mode 239 of population regulation of herbivores [2, 80]. High nutrient supply, such as in volcanic 240 or riparian soils, leads to higher quality plant forage as seen in eastern Kruger, southern 241 Serengeti and Samburu (Kenya) [71, 81]. Plants with high nutrient content and low 242 amounts of fiber can support animals of small body size that are top-down regulated. By 243 contrast, sandy soils of granitic origin are low in nutrients, and result in fibrous plants that 244 are less digestible. Such plants are eaten by large herbivores, such as elephants and 245 buffalo, which are bottom-up regulated [29, 42, 82]. This gradation forms the basis for 246 our interpretation of the different savanna systems in Africa, which we describe below.

248	Synthesis: Shifting regulation across gradients of productivity in African savannas
249	The integration of allometric differences in susceptibility to top-down and bottom-up
250	regulation along gradients of productivity enables an interpretation of the various
251	herbivore dynamics across African savannas (Figure 7). The relative importance of
252	predation, forage quality and forage abundance in regulating grazers should change under
253	different rainfall and soil fertility regimes. Figure 7 proposes how different areas in
254	Africa, often with functionally similar species, could be regulated by different
255	mechanisms along gradients of rainfall and soil fertility [27-29].
256	
257	The highest vegetative biomass occurs with high water availability (rainfall or flooding)
258	on fertile soils usually of volcanic or fluvial origin seen best in the Ruwenzori grasslands
259	(Uganda), the Nile floodplains and the flooded Boma grasslands of southwest Sudan.
260	(Figures 3b, 6a) [61, 83-85]. Under these moist fertile conditions, the quantity of food is

261 effectively unlimited and, therefore, regulates only the largest herbivores (dashed blue

line, Figure 7a). However, high rainfall causes the grass to have a large proportion of

263 poor quality stems, making digestion more difficult and reducing the overall nutritional

264 quality (dashed green line, Figure 7a). The high standing biomass under high rainfall

265 conditions also conceals predators, making small grazers more susceptible to top-down

266 effects (solid red line, Figure 7a). As a result, under high rainfall and fertile conditions,

small herbivores become regulated by predation, as in the Ruwenzori system [86],

whereas medium and large herbivores are regulated by the quality of the available forage

since the quantity is effectively limitless.

270

271	In contrast, forage quality is highest under low rainfall conditions on fertile soils, such as
272	on the open treeless Serengeti plains and the saline soils of Etosha. Here, the low grass
273	biomass exposes predators making them visible to herbivores, and results in minimal top-
274	down regulation. Instead, the quantity of forage regulates the number of medium and
275	large grazers, with predation accounting for most of the mortality in only the smallest size
276	classes (Figure 7b) [41]. The quality of the grass has little impact on regulating
277	populations, because nutritious forage is abundant and readily available.
278	
279	The nutritional quality of grass is lowest in high rainfall areas with infertile soils and,
280	therefore, becomes more important in regulating herbivore populations (Figure 7c). Such
281	areas can be found in the broad-leaved miombo woodlands of the south and central
282	African plateau, such as Selous (Tanzania), Hwange (Zimbabwe), Chobe (Botswana) and
283	Kafue (Zambia), the coastal sand dune savannas of St. Lucia (South Africa) and the
284	Maputu Elephant Reserve (Mozambique), and the moist savannas of Garamba (DRC) and
285	Mole (Ghana) [29, 87-89]. In these communities, there is a high biomass of woody
286	vegetation but also of tall grasses of lower nutritional value, which support fewer
287	herbivores and predators than areas with fertile soils. Predation is expected to be high
288	because the high grass biomass conceals predators, making them more effective [66, 68].
289	The requirements of small herbivores for high forage quality combined with their
290	susceptibility to predation limit their abundance in these areas (Figure 7c). The dominant
291	grazers are elephants [29, 89]. Even mega-herbivores might not be able to escape the
292	limitations of quality because the methane production associated with ingesting a lot of
293	low quality food reduces their capacity to absorb nutrients [39].

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295	Semi-arid systems with infertile soils, such as Tsavo and Samburu (Kenya), and the
296	Kgalagadi Transfrontier (South Africa and Botswana), tend to have sparse and
297	nutritionally poor plants, supporting lower herbivore densities and, as a result, fewer
298	predators [29, 90, 91]. These conditions are more extreme in desert environments, such
299	as the stony plains of the Jiddah (Oman) home of the Arabian oryx (Oryx leucoryx) [92],
300	and the Ouadi Achim Faunal Reserve (Chad) where the scimitar-horned oryx (O.
301	dammah) used to live [93], and the Skeleton Coast (Namibia). Under these conditions,
302	predation does not limit herbivore abundance. Herbivores are regulated by forage quality
303	(especially for smaller herbivores) and forage quantity (for larger animals) (Figure 7d).
304	

305 **Conclusions and future directions**

306 Predation and competition for resources interact synergistically rather than operate 307 independently [94]. Reciprocal, indirect, additive and interaction modifying relationships 308 shape this synergism to explain functional differences between ecosystems. In essence, 309 the interplay between: (i) the availability of limited abiotic resources (such as nutrients 310 and rainfall) that determine the quality and quantity of primary production; (ii) the 311 evolutionary tradeoffs related to body size (including predation sensitivity, digestive 312 capacity and metabolic requirements); (iii) adaptive behaviors (such as migration or 313 group vigilance), which enable primary consumers to escape regulation; and (iv) the 314 extent and frequency of disturbances (such as fires, storms, extreme temperatures, salinity 315 shifts, scouring, etc.) are processes affecting how predation and competition collectively 316 structure communities. This conceptual structure yields testable predictions for how

317 global environmental changes might affect the distribution of different sized herbivores 318 and potential regime shifts in ecosystem dynamics [95]. For example, changes in rainfall 319 due to global warming could shift the importance of food and predation in regulating 320 herbivore populations, such that decreasing rainfall would push an ecosystem along the x-321 axis in Figure 7 from panel a to b, or c to d. The evolutionary role of early hunter-322 gatherer humans in regulating herbivores as predators and as agents of disturbance fits the 323 framework of Figure 6. However, modern humans have escaped from factors regulating 324 their population density, which destabilizes framework.

325

Future research should test the predictions of Figure 7. More data are required to resolve the consequences of predation. Specifically, the analysis of herbivore carcasses suggests small prey are prone to many predators (i.e. size-nested predation, Figure 2a) [3], but this is not supported by data on carnviore diets which suggest predation is size-partitioned (Figure 2b) [4]. In addition, an evaluation of mortality in juvenile age classes might show that predation by a single predator, with low capture rates, could still impose strong population regulation, especially for larger species.

333

In summary, we propose that abiotic factors determine the importance of predation,

forage quality and forage abundance in regulating herbivores of different sizes (Figure 7)

and this alters the relative strength of the connections between biotic and abiotic

337 components in ecosystems (Figure 6a). The availability of key environmental resources

has profound consequences for herbivore regulation and ecosystem dynamics by

339 simultaneously affecting multiple top-down and bottom-up processes. The different

herbivore dynamics of the many savanna systems of Africa can be understood in the
context of this framework. These concepts could help our understanding of other
ecosystems where strong abiotic gradients influence the shape of the community (such as
salinity and desiccation in intertidal ecosystems, dissolved oxygen and opacity in aquatic
ecosystems, or body mass and predation risk in avian communities). The strength of this
framework is that it captures how environmental gradients can switch top-down and
bottom-up processes that regulate animal abundance.

347

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358 **References**

- Harris, G. *et al.* (2009) Global decline in aggregated migrations of large terrestrial mammals. *Endang. Sp. Res.* 7, 55-76
- Schmitz, O.J. (2008) Herbivory from individuals to ecosystems. *Annu. Rev. Ecol. Evol. Syst.* 39, 133-152
- 363 3. Sinclair, A.R.E. *et al.* (2003) Patterns of predation in a diverse predator-prey system.
 364 *Nature* 425, 288-290
- 365 4. Owen-Smith, N. and Mills, M.G.L. (2008) Predator-prey size relationships in an
 366 African large-mammal food web. *J. Anim. Ecol.* 77, 173-183
- 5. Estes, J.A. and Duggins, D.O. (1995) Sea Otters and kelp forests in Alaska Generality and variation in a community ecological paradigm. *Ecol. Monogr.* 65, 75100
- 6. Coe, M.J. *et al.* (1976) Biomass and production of large African herbivores in relation
 to rainfall and primary production. *Oecologia* 22, 341-354
- 372 7. Polis, G.A. (1999) Why are parts of the world green? Multiple factors control
 373 productivity and the distribution of biomass. *Oikos* 86, 3-15
- McNaughton, S.J. *et al.* (1989) Ecosystem-level patterns of primary productivity and herbivory in terrestrial habitats. *Nature* 341, 142-144
- 376 9. East, R. (1984) Rainfall, soil nutrient status and biomass of large African savanna
 377 mammals. *Afr. J. Ecol.* 22, 245-270
- 378 10. Knight, T.M. et al. (2005) Trophic cascades across ecosystems. Nature 437, 880-883
- 379 11. Hairston, N.G. *et al.* (1960) Community structure, population control, and
 380 competition. *Am. Nat.* 94, 421-425
- 12. Terborgh, J. *et al.* (2006) Vegetation dynamics of predator-free land-bridge islands. *J. Ecol.* 94, 253-263
- 383 13. Cohen, J.E. *et al.* (1993) Body sizes of animal predators and animal prey in food
 384 webs. *J. Anim. Ecol.* 62, 67-78
- 14. Rooney, N. *et al.* (2008) A landscape theory for food web architecture. *Ecol. Lett.* 11,
 867-881
- 15. Radloff, F.G.T. and Du Toit, J.T. (2004) Large predators and their prey in a southern
 African savanna: a predator's size determines its prey size range. *J. Anim. Ecol.* 73,
 410-423
- 390 16. Creel, S. (2001) Four factors modifying the effect of competition on carnivore
 391 population dynamics as illustrated by African wild dogs. *Conserv. Biol.* 15, 271-274
- 392 17. Palomares, F. and Caro, T.M. (1999) Interspecific killing among mammalian
 393 carnivores. *Am. Nat.* 153, 492-508

- 394 18. Scheel, D. and Packer, C. (1991) Group hunting behaviour of lions: a search for
 395 cooperation. *Anim. Behav.* 41, 697-709
- Hayward, M.W. and Kerley, G.I.H. (2005) Prey preferences of the lion (*Panthera leo*). *Journal of Zoology* 267, 309-322
- 20. Peters, R.H. (1983) *The Ecological Implications of Body Size*, Cambridge University
 Press
- 400 21. Owen-Smith, N. (1988) Megaherbivores: The Influence of Very Large Body Size On
 401 Ecology, Cambridge University Press
- 402 22. Botkin, D.B. *et al.* (1981) How ecosystem processes are linked to large mammal
 403 population dynamics. In *Dynamics of Large Mammal Populations* (Fowler, C.W. and
 404 Smith, T.D., eds), pp. 373-387, John Wiley & Sons
- 405 23. Scott, R.M. (1962) Exchangeable bases of mature, well-drained soils in relation to
 406 rainfall in East Africa. J. Soil Sci. 13, 1-9
- 407 24. Hook, P.B. and Burke, I.C. (2000) Biogeochemistry in a shortgrass landscape:
 408 Control by topography, soil texture, and microclimate. *Ecology* 81, 2686-2703
- 409 25. Jacobs, S.M. *et al.* (2007) Nutrient vectors and riparian processing: A review with
 410 special reference to African semiarid Savanna ecosystems. *Ecosystems* 10, 1231-1249
- 411 26. Gagnon, M. and Chew, A.E. (2000) Dietary preferences in extant African bovidae. *J.*412 *Mammal.* 81, 490-511
- 413 27. Fritz, H. *et al.* (2002) Megaherbivores influence trophic guilds structure in African
 414 ungulate communities. *Oecologia* 131, 620-625
- 415 28. Fritz, H. and Duncan, P. (1994) On the carrying-capacity for large ungulates of
 416 African savanna ecosystems. *Proc. R. Soc. B* 256, 77-82
- 417 29. Bell, R., H, V (1982) The effect of soil nutrient availability on community structure in
 418 African ecosystems. In *Ecology of Tropical Savannas* (Huntley, B.J. and Walker,
 419 B.H., eds), pp. 193-216, Springer-Verlag
- 30. Treydte, A.C. *et al.* (2007) Trees improve grass quality for herbivores in African
 savannas. *Perspect. Plant Ecol. Evol. Syst.* 8, 197-205
- 422 31. Oksanen, L. *et al.* (1981) Exploitation ecosystems in gradients of primary
 423 productivity. *Am. Nat.* 118, 240-261
- 424 32. Fryxell, J.M. (1991) Forage quality and aggregation by large herbivores. *Am. Nat.*425 138, 478-498
- 426 33. Prins, H.T. and Olff, H. (1998) Species richness of African grazer assemblages:
 427 towards a functional explanation. In *Dynamics of Tropical Communities* (Newbery,
 428 D.M. *et al.*, eds), pp. 449-490, Blackwell Science
- 429 34. Mduma, S.A.R. *et al.* (1999) Food regulates the Serengeti wildebeest: A 40-year
 430 record. *J. Anim. Ecol.* 68, 1101-1122
- 431 35. Olff, H. *et al.* (2002) Global environmental controls of diversity in large herbivores.
 432 *Nature* 415, 901-904

- 433 36. McNaughton, S.J. *et al.* (1985) Silica as a defense against herbivory and a growth
 434 promotor in African grasses. *Ecology* 66, 528-535
- 435 37. Gordon, I.J. and Illius, A.W. (1996) The nutritional ecology of African ruminants: A
 436 reinterpretation. *J. Anim. Ecol.* 65, 18-28
- 437 38. Sues, H.D. and Reisz, R.R. (1998) Origins and early evolution of herbivory in
 438 tetrapods. *Trends Ecol. Evol.* 13, 141-145
- 439 39. Clauss, M. and Hummel, J. (2005) The digestive performance of mammalian
 440 herbivores: why big may not be that much better. *Mammal. Rev.* 35, 174-187
- 40. Wilmshurst, J.F. *et al.* (2000) The allometry of patch selection in ruminants. *Proc. R. Soc. B* 267, 345-349
- 443 41. Jarman, P.J. (1974) Social-organization of antelope in relation to their ecology.
 444 *Behaviour* 48, 215-&
- 445 42. Codron, J. *et al.* (2006) Elephant (*Loxodonta africana*) diets in Kruger National Park,
 446 South Africa: Spatial and landscape differences. *J. Mammal.* 87, 27-34
- 447 43. van de Koppel, J. *et al.* (2002) Spatial heterogeneity and irreversible vegetation
 448 change in semiarid grazing systems. *Am. Nat.* 159, 209-218
- 449 44. Scheffer, M. and Carpenter, S.R. (2003) Catastrophic regime shifts in ecosystems:
 450 linking theory to observation. *Trends Ecol. Evol.* 18, 648-656
- 45. McNaughton, S.J. *et al.* (1988) Large mammals and process dynamics in African
 ecosystems. *Bioscience* 38, 794-800
- 46. Archibald, S. *et al.* (2005) Shaping the landscape: Fire-grazer interactions in an
 African savanna. *Ecol. Appl.* 15, 96-109
- 47. Hebblewhite, M. (2005) Predation by wolves interacts with the North Pacific
 456 Oscillation (NPO) on a western North American elk population. *J. Anim. Ecol.* 74,
 457 226-233
- 458 48. Arsenault, R. and Owen-Smith, N. (2002) Facilitation versus competition in grazing
 459 herbivore assemblages. *Oikos* 97, 318
- 460 49. Vesey-FitzGerald, D.F. (1960) Grazing succession among East African game
 461 animals. J. Mammal. 41, 161-172
- 462 50. Cromsigt, J. and Olff, H. (2008) Dynamics of grazing lawn formation: an
 463 experimental test of the role of scale-dependent processes. *Oikos* 117, 1444-1452
- 464 51. Waldram, M.S. *et al.* (2008) Ecological engineering by a mega-grazer: White Rhino
 465 impacts on a South African savanna. *Ecosystems* 11, 101-112
- 466 52. McNaughton, S.J. (1988) Mineral nutrition and spatial concentrations of African
 467 ungulates. *Nature* 334, 342-345
- 468 53. Anderson, T.M. *et al.* (2009) Landscape-scale analyses suggest both nutrient and anti 469 predator advantages to Serengeti hotspots. *Ecology* revision invited
- 470 54. Kauffman, J.B. *et al.* (1994) Relationships of fire, biomass and nutrient dynamics
 471 along a vegetation gradient in the Brazilian cerrado. *J. Ecol.* 82, 519-531

472 55. Dublin, H.T. et al. (1990) Elephants and fire as causes of multiple stable states in the 473 Serengeti-Mara Tanzania woodlands. J. Anim. Ecol. 59, 1147-1164 474 56. Bond, W.J. and Keeley, J.E. (2005) Fire as a global 'herbivore': the ecology and 475 evolution of flammable ecosystems. Trends Ecol. Evol. 20, 387-394 476 57. Belsky, A.J. (1992) Effects of grazing competition disturbance and fire on species 477 composition and diversity in grassland communities. J. Veg. Sci. 3, 187-200 478 58. Sankaran, M. et al. (2005) Determinants of woody cover in African savannas. Nature 479 438, 846-849 480 59. Higgins, S.I. et al. (2007) Effects of four decades of fire manipulation on woody 481 vegetation structure in savanna. Ecology 88, 1119-1125 482 60. Olff, H. et al. (2009) Parallel ecological networks in ecosystems. Phil. Trans. R. Soc. 483 *B* 364, 1755-1779 484 61. Anderson, T.M. et al. (2007) Forage nutritive quality in the Serengeti ecosystem: the 485 roles of fire and herbivory. Am. Nat. 170, 343-357 486 62. Anderson, T.M. (2008) Plant compositional change over time increases with rainfall 487 in Serengeti grasslands. Oikos 117, 675-682 488 63. Bond, W.J. (2008) What limits trees in C-4 grasslands and savannas? Annu. Rev. 489 Ecol. Evol. Syst. 39, 641-659 490 64. Owen-Smith, N. et al. (2005) Correlates of survival rates for 10 African ungulate 491 populations: density, rainfall and predation. J. Anim. Ecol. 74, 774-788 492 65. Hebblewhite, M. et al. (2005) Spatial decomposition of predation risk using resource 493 selection functions: an example in a wolf-elk predator-prey system. Oikos 111, 101-494 111 495 66. Hopcraft, J.G.C. et al. (2005) Planning for success: Serengeti lions seek prev 496 accessibility rather than abundance. J. Anim. Ecol. 74, 559-566 497 67. Kauffman, M.J. et al. (2007) Landscape heterogeneity shapes predation in a newly 498 restored predator-prey system. Ecol. Lett. 10, 690-700 499 68. Balme, G. et al. (2007) Feeding habitat selection by hunting leopards Panthera pardus 500 in a woodland savanna: prey catchability versus abundance. Anim. Behav. 74, 589-501 598 502 69. Riginos, C. and Grace, J.B. (2008) Savanna tree density, herbivores, and the 503 herbaceous community: Bottom-up vs. top-down effects. Ecology 89, 2228-2238 504 70. Wang, G.M. et al. (2006) Spatial and temporal variability modify density dependence 505 in populations of large herbivores. Ecology 87, 95-102 506 71. Venter, F.J. et al. (2003) The abiotic template and its associated vegetation pattern. In 507 The Kruger Experience: Ecology and Management of Savanna Heterogeneity (du 508 Toit, J.T. et al., eds), pp. 83-129, Island Press 509 72. Cromsigt, J. and Olff, H. (2006) Resource partitioning among savanna grazers 510 mediated by local heterogeneity: An experimental approach. Ecology 87, 1532-1541

511 73. Fryxell, J.M. and Sinclair, A.R.E. (1988) Causes and consequences of migration by 512 large herbivores. Trends Ecol. Evol. 3, 237-241 513 74. Fryxell, J.M. et al. (1988) Why are migratory ungulates so abundant? Am. Nat. 131, 514 781-798 515 75. Gasaway, W.C. et al. (1996) Persistent low densities of plains ungulates in Etosha 516 National Park, Namibia: Testing the food-regulating hypothesis. Can. J. Zool. 74, 517 1556-1572 518 76. Sinclair, A.R.E. et al. Trophic cascades in african savanna: Serengeti as a case study 519 (Estes, J.A. and Terborgh, J., eds), Island Press (in press) 520 77. Tambling, C.J. and Du Toit, J.T. (2005) Modelling wildebeest population dynamics: 521 implications of predation and harvesting in a closed system. J. Appl. Ecol. 42, 431-522 441 523 78. Cromsigt, J.P.G.M. et al. (2009) Habitat heterogeneity as a driver of ungulate 524 diversity and distribution patterns: interaction of body mass and digestive strategy. 525 Divers. Distrib. 15, 513-522 526 79. Valeix, M. et al. (2009) Behavioral adjustments of African herbivores to predation 527 risk by lions: Spatiotemporal variations influence habitat use. Ecology 90, 23-30 528 80. Owen-Smith, N. and Mills, M.G.L. (2008) Shifting prey selection generates 529 contrasting herbivore dynamics within a large-mammal predator-prey web. *Ecology* 530 89, 1120-1133 531 81. Eckhardt, H.C. et al. (2000) Trends in woody vegetation cover in the Kruger National 532 Park, South Africa, between 1940 and 1998. Afr. J. Ecol. 38, 108-115 533 82. Bell, R.H.V. (1970) The use of herb layer by grazing ungulates in the Serengeti. In 534 Animal Populations in Relation to Their Food Resources (Watson, A., ed), pp. 111 -535 124, Blackwell 536 83. Eltringham, S.K. and Woodford, M.H. (1973) The numbers and distribution of 537 buffalo in the Ruwenzori National Park, Uganda. Afr. J. Ecol. 11, 151-164 538 84. Fryxell, J.M. and Sinclair, A.R.E. (1988) Seasonal migration by white-eared kob in 539 relation to resources. Afr. J. Ecol. 26, 17-31 540 85. Howell, P. et al. (1988) The Jonglei Canal: Impact and Opportunity, Cambridge 541 University Press 542 86. Yoaciel, S.M. (1981) Changes in the populations of large herbivores and in the 543 vegetation community in Mweya Peninsula, Rwenzori National Park, Uganda. Afr. J. 544 *Ecol.* 19, 303-312 545 87. Ellenbroek, G.A. (1987) Ecology and productivity of an African wetland system: the 546 Kafue Flats, Zambia, Dr. W. Junk 547 88. Mosugelo, D.K. et al. (2002) Vegetation changes during a 36-year period in northern 548 Chobe National Park, Botswana. Afr. J. Ecol. 40, 232-240

- 549 89. Valeix, M. *et al.* (2008) Fluctuations in abundance of large herbivore populations:
 550 insights into the influence of dry season rainfall and elephant numbers from long551 lterm data. *Anim. Conserv.* 11, 391-400
- 90. Parker, I.S.C. (1983) The Tsavo story: an ecological case history. In *Management of large mammals in African conservation areas* (Owen-Smith, N., ed), pp. 37-49,
 Haum
- 555 91. Bergstrom, R. and Skarpe, C. (1999) The abundance of large wild herbivores in a 556 semi-arid savanna in relation to season, pans and livestock. *Afr. J. Ecol.* 37, 12-26
- 557 92. Stanley-Price, M.R. (1989) *Animal Re-introductions: The Arabian Oryx in Oman*,
 558 Cambridge University Press
- 93. Newby, J.E. (1974) The ecological resources of the Ouadi-Rime-Ouadi Achim Faunal
 Reserve, Chad. In *Direction des Parcs Nationaux et Reserves de Faune*, F.A.O.
- 561 94. Chesson, P. and Kuang, J.J. (2008) The interaction between predation and
 562 competition. *Nature* 456, 235-238
- 563 95. Tylianakis, J.M. *et al.* (2008) Global change and species interactions in terrestrial
 564 ecosystems. *Ecol. Lett.* 11, 1351-1363
- 565 96. Fryxell, J.M. *et al.* (2007) Group formation stabilizes predator-prey dynamics. *Nature*566 449, 1041-1044
- 567 97. Ostrowski, S. *et al.* (2006) Physiological acclimation of a desert antelope, Arabian
 568 oryx (*Oryx leucoryx*), to long-term food and water restriction. *J. Comp. Physiol. B*569 176, 191-201
- 98. van Soest, P.J. (1996) Allometry and ecology of feeding behavior and digestive
 capacity in herbivores: A review. *Zoo Biol.* 15, 455-479
- 572 99. Blair, J.M. (1997) Fire, N availability, and plant response in grasslands: A test of the
 573 transient maxima hypothesis. *Ecology* 78, 2359-2368
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575

577 Box 1. Adaptive responses to regulation: migrations, crypsis and vigilance

578

5/8	
579	Large herbivores are occasionally found in areas of low plant biomass while small
580	herbivores are sometimes found in areas of high plant biomass, which is contrary to the
581	expectations of Figure 7 (main text). The realized niche of many herbivores differs from
582	expected because they escape regulation through behavioral and physiological
583	adaptations selected through evolutionary time [26].
584	
585	For example, migrations enable a population to escape the limitations of both low forage
586	biomass and predation. By sequentially moving between the best available food patches
587	migrants are in essence increasing the total available biomass of highest quality food,
588	without exhausting the overall food supply. Small resident populations of wildebeest in
589	Pilanesberg, Kruger and Serengeti are regulated by predation [64, 77], whereas the large
590	migratory wildebeest population in the Serengeti (~1.2 million) are regulated by the
591	availability of sufficient forage especially during the dry season when food is most
592	limited [34]. Other resident herbivore species in Kruger (e.g. roan, Hippotragus equinus)
593	and Serengeti become more predator regulated in areas they share with migrant
594	populations, because the seasonal influx of migratory prey supports more predators that
595	tend to supplement their diet with resident prey [80]. As a result, the abundances of
596	migrants versus residents can differ by orders of magnitude [73, 74, 76]. In
597	circumstances where migratory routes are blocked the population declines rapidly [1] and

598 becomes predator regulated. Migrations explain how very large herbivores can persist in

599 low biomass areas, such as the paradox of mammoths which occupied the low biomass

600	habitats of the tundra. Other behavioral adaptations to predation include group dilution
601	and group vigilance where individuals decrease their probability of being caught [96].
602	
603	Avoiding predators through habitat selection or crypsis requires physiological
604	adaptations. For instance, addax (Addax nasomaculatus), gemsbok (Oryx gazelle) and
605	Arabian oryx avoid predators by living in very arid habitats, however they require
606	complex adaptations for water retention and cooling [97]. Digestive adaptations are
607	required especially by small energetically constrained herbivores relying on crypsis in
608	thick low-quality vegetation [98]. In both circumstances, populations escape predation
609	but become regulated at much lower densities by forage quantity and quality,

610 respectively.

- Figure 1. Several factors are responsible for regulating herbivore abundances across
 ecosystems. In this classic food chain approach, the abundance of herbivores is
- 613 regulated by top-down processes, such as predation, and by bottom-up processes
- 614 through primary production (arrows indicate the direction of influence).
- 615 Herbivores (like carnivores) are considered as a single unit, despite showing strong 616 functional divergences based on body size. Primary production in this framework
- 617 is a general term that fails to distinguish the differences between the quality,
- 618 quantity, and structure of a plant community.
- 619
- 620 Figure 2. The degree of herbivore mortality due to predation depends on the predators' 621 diet selection. (a) If large predators are opportunists and consume prey of all sizes, 622 while small predators only kill small prey, then the prey base of small predators is 623 nested within that of large predators (size-nested predation). The cumulative 624 mortality on small prey is much greater than large prey because they are exposed to 625 more predators. (b) If predators are selective, and only consume prey of a specific 626 size classes (size-partitioned predation), then large predators do not supplement 627 their diet with small prey. When predation is size-partitioned as opposed to sizenested, the cumulative mortality due to predation on small prey is much less, 628 629 whereas large prey are killed more often.
- 630 631 Figure 3. The relative importance of predation in regulating herbivore populations 632 depends on the body size of predators and their degree of specialization for certain 633 prey. (a) If large carnivores only eat large prey, and small carnivores only eat small 634 prey (solid vertical lines), the prey base is partitioned. If large carnivores eat both 635 large and small prey (solid and dashed vertical lines), the prey base of smaller 636 carnivores is nested within that of larger carnivores. Large carnivores dominate 637 small carnivores and reduce their efficiency (solid horizontal line). (b) When predation is nested, small prev are exposed to more predator species and become 638 639 increasingly predator regulated, as in the Serengeti example (solid line). When 640 predation is partitioned, large prey suffer greater predation than do small prey 641 because large predators do not supplement their diet with small prey, as in the 642 Kruger example (dashed line). Data for Serengeti and Kruger reproduced with 643 permission from Refs. [3, 4, 64]. (Abbreviations: B = African Buffalo, E = 644 Elephant, G = Giraffe, H = Hippo, I = Impala, O = Oribi, R = Black Rhino, T = Topi, W = resident Wildebeest, Z = resident Zebra) 645
- 646 647
- 648 Figure 4. The quality versus quantity of primary production regulates large herbivores 649 differently from small herbivores. (a) Physical and environmental gradients have 650 direct and indirect affects (i.e. between non-adjacent levels) on the plant community structure and on the quality and quantity of primary production. (b) For 651 example, the quantity of plant biomass is positively related to increasing rainfall 652 and soil fertility, whereas the digestible quality of the plant declines with increasing 653 654 rainfall. (c) Large herbivores consume greater quantities of lower quality food, 655 whereas small herbivores consume less food of higher quality because they are constrained by their high metabolism and limited digestive capacity. (d) Therefore, 656

large herbivores, such as elephant, tend to be regulated by food abundance (dashed
blue line), whereas smaller herbivores, such as wildebeest, are regulated by food
quality (dotted green line). The smallest herbivores, such as oribi, are mainly
predator regulated (solid red line). Reproduced with permission from Refs. [35] (b)
and [21] (c).

662 663

664 Figure 5. Abiotic and biotic disturbances, such as fires or intense grazing, alter the primary production and the plant community structure of a landscape which, in 665 666 turn, affects the abundance of different sized herbivores. (a) Disturbances can have reciprocal effects (double arrows) on primary production. For example, herbivores 667 can reduce the biomass of grass, which reduces the probability of fire; but, 668 conversely, fires removes grass necromass which stimulates re-growth of high 669 670 quality shoots that are preferred by herbivores. Interaction-modifying relationships (dotted arrow) alter the effects of a disturbance, such as large amounts of 671 672 flammable biomass altering the severity of a fire. Positive feed-back loops have 673 additive effects, such as grass biomass increasing the probability of a fire, which removes trees, and provides grass with a competitive advantage. (b) Biotic 674 675 disturbances, such as grazing lawns created by white rhino, modify the quality and 676 quantity of vegetation over time by altering the competitive balance between grazing tolerant and grazing intolerant grass species (described by Refs. [46, 50, 677 678 51]). (c) Abiotic disturbances, such as fires, alter the short-term abundance and 679 nutritional quality of the grasses available to herbivores by removing senescent vegetation and stimulating nutrient-rich re-growth (described by Refs [56, 61, 99]). 680 681 Long-term species succession, could reverse this trend.

682 683

684 Figure 6. An ecosystem template of the macroecological processes determining the 685 abundance and distribution of herbivores. (a) Underlying environmental gradients simultaneously affect both the quality and quantity of forage available to herbivores 686 as well as the efficiency of predators at capturing prey. Thus, bottom-up and top-687 688 down processes are not independent. For example, (b) lions select areas with denser vegetation, and (c) areas that are closer to rivers for hunting more often than 689 690 expected, based on the availability of these resources across the landscape. 691 Therefore, the plant community structure (such as % tree cover) and topographic 692 features (such as rivers) contribute to the predation risk for herbivores, while simultaneously influencing the quality and quantity of forage available to them 693 694 (data from Ref. [66]).

695 696

Figure 7. Predictions of the relative importance of predation (solid red line), food quality
(dotted green line) and food abundance (dashed blue line) in regulating herbivores
of increasing body mass across rainfall and soil fertility gradients, assuming all else
is equal (e.g. availability of drinking water, size and isolation of protected area,
etc.). We consider herbivores between 10kg to 1000kg, with a major portion of
grass in their diet [33]. (a) High rainfall and soil nutrients. Food abundance

- 703 regulates large herbivores, food quality regulates medium sized herbivores and 704 predation regulates small herbivores. (b) Low rainfall, high soil nutrients. Food 705 quality does not regulate. Food abundance regulates large and medium sized 706 herbivores, predation regulates small ones. (c) High rainfall, low soil nutrients. 707 Food quality regulates all herbivores, with predation acting synergistically at small 708 size, and food abundance at very large size. (d) Low rainfall and soil nutrients. 709 Predation is not regulating, food quality acts at small size and food abundance at 710 medium and large sizes. For more details see main text. If rainfall patterns change 711 due to global warming, herbivore regulation within a savanna ecosystem is 712 predicted to shift from panel a to b, or c to d, or visa versa. 713
- 714
- 715

Figure 1

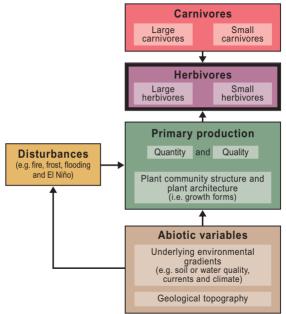
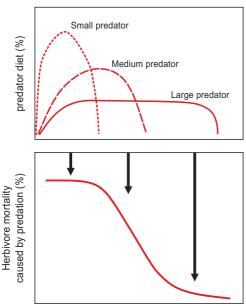
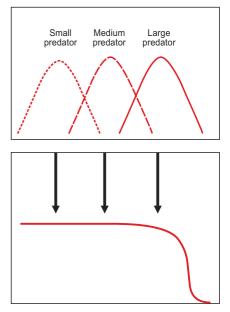


Figure 2

(a) Size-nested predation



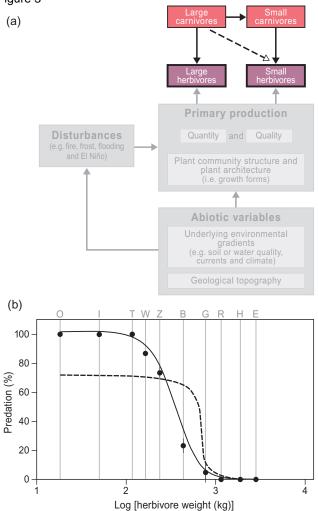
(b) Size-partitioned predation

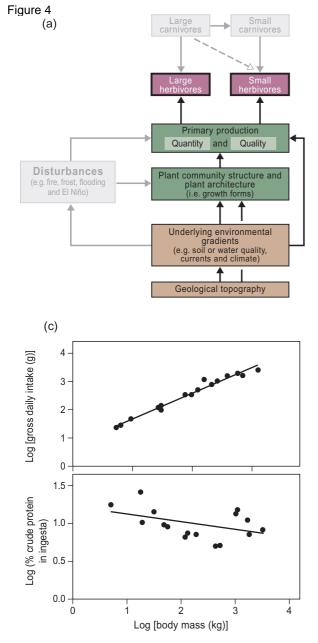


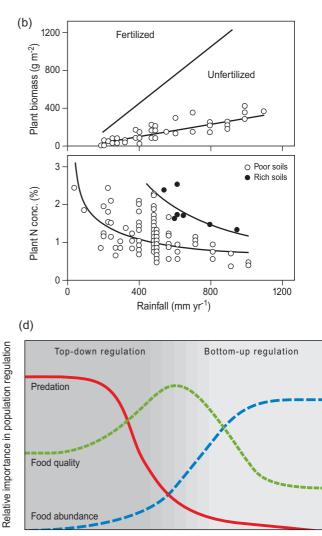
Prey body mass

Prey body mass



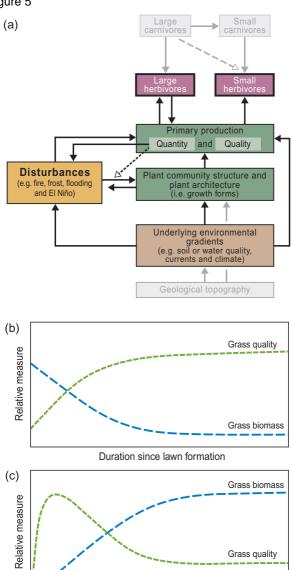






Oribi Impala Wildebeest Buffalo Hippo Duiker Nyala Zebra Elephant Klipspringer Th. gazelle Waterbuck Dikdik Greater kudu







Grass quality

Figure 6

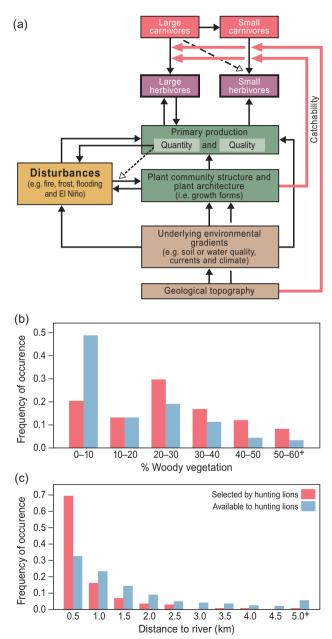
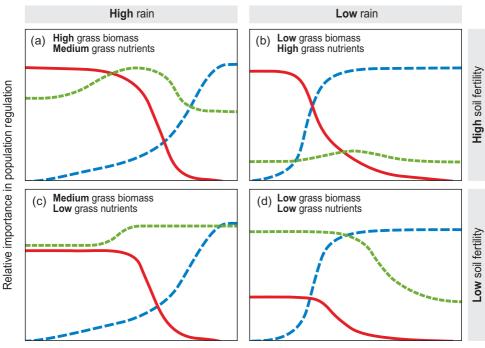


Figure 7



Predation
 Food quality
 Food abundance

Body size